Abstract: A comparison of benthic polychaete communities off the Sand Island Wastewater Outfall was undertaken to recognize organic enrichment indicator species for Hawaiian waters. Primary-treatment sewage is discharged off the south shore of O`ahu at 70 m depth. A historical data set spanning 9 yr for seven sites at 70 m and two recent studies at 20, 50, and 100 m depths were analyzed. Geochemical data did not support the assumption that the outfall is an important source of organic enrichment in nutrient-poor sandy sediments within oligotrophic tropical waters. Five polychaete species, however, appeared particularly sensitive, positively or negatively, to environmental conditions near the outfall. Neanthes arenaceodentata (Nereididae) and Ophryotrocha adherens (Dorvilleidae) have been dominant at sites within the outfall’s zone of initial dilution (ZID). Since 1993, N. arenaceodentata has virtually disappeared, and O. adherens concurrently became abundant and continued to flourish at ZID sites. Well-known indicators within the Capitella capitata complex (Capitellidae) were present at ZID and control (far field) sites though their ZID abundance was greater. Two sabellids, Euchone sp. B and Augeneriella dubia were inversely distributed, the smaller Euchone sp. B at far field sites and larger A. dubia within ZID stations. The former was most likely restricted to a greater proportion of fine sediment particles at two far field sites. The most abundant and widespread polychaete off O`ahu’s south shore was Pionosyllis heterocirrata (Syllidae), which does not seem to represent a sensitive indicator species. Ophryotrocha adherens was the most abundant indicator species within the ZID; P. heterocirrata was the most ubiquitous species at all sites and should always be expected in these sediments. Traditional measurements of numerical abundance, species richness, and diversity (H’) have not shown a clear distinction between ZID and far field sites in annual analyses. An examination of composited data over an 11-yr period does support such a distinction. Multidimensional scaling (MDS) analyses clearly delineate different assemblages. We suggest that MDS analyses are sensitive to the community differences present near the outfall. The ZID community is clearly contained within the Environmental Protection Agency-approved ZID boundary. Because each ZID and far field site supports a diverse and coarsely similar polychaete fauna, no pollution level effects seem to be present.
poses. Benthic fauna are regularly used because they often represent a substantial portion of the biomass or supply much of the food to the more visible and sometimes commercially valuable water column inhabitants. The methods and models chosen for a particular study often depend upon the investigator's knowledge of the local habitat, trained personnel available for field and laboratory work, or government regulations and approved protocols.

The University of Hawai'i Water Resources Research Center began monitoring environmental conditions surrounding the Sand Island Wastewater Outfall in 1986. This study presents a synthesis of data compiled from the Water Resources Research Center's annual reports from 1990 to 2000 when sampling and reporting practices were standardized as presented in Nelson et al. (1992a,b, 1994, 1995, 1996, 1997) and Swartz et al. (1998, 1999). The program has been designed to examine the abundance of expected polychaete species and other faunal components within the area defined as the zone of initial dilution (ZID) when compared with similar control sites (far field). Polychaetes are particularly useful because of the relative ease of identification, a trophic position that may respond to the organic enrichment produced by the outfall, their ubiquity in Hawaiian sediments, and densities that often approach 100,000 individuals per square meter. The binary sampling design, ZID versus non-ZID, was chosen to provide ready statistical comparison and to develop a reliable characterization of the ZID-associated community in preparation for more regional and gradient-based studies currently under way.

We base our fundamental assessments on one of the most broadly applied models of biological response to organic enrichment used in many studies from North America, northern Europe, Australia, and India: the Pearson-Rosenberg model (Pearson and Rosenberg 1978). In essence, the benthic community composition is measured along a gradient from a source of organic enrichment to areas outside the zone of influence. The model predicts an increase in species richness and evenness of abundance with increasing distance from the input until the usual abundance, evenness, and species richness for indigeneous balanced communities are reached. Numerically dominant polychaete species play a major role in shaping benthic communities and are regarded as indicators of environmental quality (Pocklington and Wells 1992). Changes in species richness, diversity, community evenness, and movement within dominance rankings suggest biological responses to environmental changes and may be reason for concern unless the change can be attributed to a known and benign factor. Graded changes may reflect the dynamic nature of benthic assemblages and provide valuable, time-sensitive baseline data. Rapid and/or inexplicable changes may suggest that a closer examination is called for and that continued monitoring is necessary. In addition to community-level analyses, we examine the 10 numerically dominant species that typically represent 70 to 90% of collected specimens. By characterizing the life histories of these species and the fluctuations in their abundance from 1990 to the present we hope to gain a better understanding of these benthic assemblages and assess the impacts of organic enrichment by comparing sites within the ZID boundary with those outside it.

Recognizing a species as a "pollution indicator" can have several meanings. Various species have been labeled as indicators for reasons not necessarily consistent among impact studies (Jones and Kaly 1996). Every region is likely to present a different series of positive and negative organic enrichment indicators. Positive indicators may be ecological opportunists whose feeding style, reproduction, and growth allow the population to increase concomitantly with organic input over a relatively short period of time (Davis 1987, Bailey-Brock 1996). Negative indicators are those sensitive species that die directly as a result of conditions related to habitat alteration or are excluded by competitive interactions within the altered habitat.

When comparing invertebrate assemblages between any two sites, within the context of the Pearson-Rosenberg model, it is important to consider influences besides organic en-
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Airport Area of Detail Honolulu

Figure 1. Study area and sample sites in relation to the O'ahu shoreline and 70-m isobath. Open circles denote 70 m sites sampled from 1990 to 1998. Solid circles indicate 1999 and 2000 study sites at 20 m (C), 50 m (D), and 100 m (E) depths. WWTP, Sand Island Wastewater Treatment Plant. The border surrounding the outfall’s seaward end marks the ZID boundary.

richment that can vary between sites. For benthic polychaetes, many of which are microscopic, the sediment grain size distribution is one such important consideration because space between the grains defines the physical nature of the habitat. The physical composition of sediments broadly affects the available habitat microtopography, oxygen and waste exchange, building material availability, and interstitial stability. Each species’ motility and feeding methods may be more or less successful within any particular granulometric environment. Field observations of these animals are extremely difficult to obtain, so our assumptions of motility and feeding methods are based in part on generalizations made by Fauchald and Jumars (1979), deductions based on morphology, and our brief observations of live animals in the laboratory.

Materials and Methods

Sampling Sites and Collection Procedures

The Sand Island Wastewater Treatment Plant on the south shore of O'ahu, Hawai'i, discharges 60–78 million gallons per day (2.6–3.4 m³/sec) of primary-treatment effluent at a depth of 70 m over a sandy bottom. Seven stations at approximately 70 m depth, four around the diffuser within the ZID, and three approximately 1, 1.5, and 2 km distant were monitored from 1986 to 1998 (e.g., Nelson et al. 1994). Since then 15 sites at approximately 20, 50, and 100 m depths were used instead (Swartz et al. 2001) to establish the benthic community composition surrounding the 70-m contour (Figure 1).

A single sediment core, 7.6 cm in diameter and minimally 5 cm deep, was taken from one
deployment of a modified 0.16-m$^2$ van Veen grab fitted with top-closing flaps. The core was fixed in 10% buffered formalin, processed to extract the fauna over a 0.5-mm sieve (Nelson et al. 1994), and preserved in 70% ethanol in water. Six replicate grabs were obtained in the historical study (1990–1998), and three replicates were obtained in the 1999 and 2000 collections. Polychaetes were identified and enumerated. Replicated sediment samples from each station were processed by a dry-sieve method for grain size analyses (Nelson et al. 1994). Chemical parameters of sediments were determined by U.S. Environmental Protection Agency–approved methods conducted by the City and County of Honolulu and various contract agencies as reported in Nelson et al. (1992a,b, 1994, 1995, 1996, 1997) and Swartz et al. (1998, 1999).

Statistical Analyses

Diversity was considered using the Shannon-Wiener statistic ($H'$) and Pielou's evenness of abundance across species ($J$). The $H'$ statistic is often cited without cautioning the reader that it assumes that all species are present in each sample. This is a serious drawback for our polychaete community analyses because a large percentage of enumerated species belong to only a few taxa throughout the study locale even in "healthy" far field stations. Pielou’s evenness takes on a value of -99 when all individuals in a sample belong to one species and a value of +1 when each taxon contributes one individual to the sample. The $K$-dominance curves, diversity ($H'$) and evenness ($J$) statistics, and multidimensional scaling (MDS) analyses were calculated using Plymouth Software’s Primer Version 4 (Carr 1997). $K$-dominance curves (Lambshead et al. 1983) have historically been used to help differentiate between communities dominated by one or a few species as illustrated by different curve profiles.

Minitab for Windows Release 12.1 (Minitab, Inc.) was used for tests of significance at the confidence intervals specified, and for linear regression and correlation analyses. Station data for any given year were found to be homoscedastic. Normal distribution was assumed when random observations of abundance, species richness, and derived variables ($H'$ and $J$) were analyzed with Minitab.

MDS analyses are increasingly used to directly consider benthic communities with few, if any, assumptions about the data. Krassulya’s (2001) comparison of statistical methods useful for investigating organically enriched invertebrate communities strongly supports MDS analyses as the most sensitive method of identifying community differences. Techniques such as $K$-dominance curves and comparisons of richness, abundance, and diversity as used here are robust techniques, but may only detect distinctions at extreme pollution levels. In any monitoring program it is desirable to constantly characterize the impacted areas and provide timely information about changes toward a polluted condition rather than simply affirm whether or not a drastic result has occurred.

Primer's Anosim subroutine was used to compare MDS groupings in one-way and two-way nested tests. Because of software limitations, the 1990–1998 historical data set was split in half to produce one MDS ordination for 1990–1993 data and another for 1994–1998 data. Species contributing less than 5% of total collected polychaetes were removed from both data sets. Samples were removed randomly to accommodate matrix limits of the software. Ultimately, 61 polychaete species (of 208 observed) among 120 of the 1990–1993 samples (168 total) and 120 samples from 1994–1998 (210 total) were considered in these analyses. All polychaete taxa were included in diversity, evenness, and richness values. Similarity matrices were constructed using the software defaults of Bray-Curtis similarity index, and fourth root transformation of abundance. Graphically representing the three-dimensional similarity array in two dimensions is an iterative process. Primer was therefore instructed to perform 90 random starts, of which the most accurate representation (least stress) is depicted. $K$-dominance and MDS analyses were conducted on both raw numerical abundance and percentage abundance within samples. Because random data reduction was required,
the nonparametric analyses were performed three times. Results from raw data, percentage values, and repeated randomizations were nearly identical. All abundance data presented here are in terms of percentage abundance within a single sample unless otherwise stated.

**Indicator Species**

**Dorvilleidae**  
*Ophryotrocha adherens* Paavo, Bailey-Brock & Akesson, 2000

These are minute (1.2 mm long) jawed dorvilleids composed of 17–20 segments. Each segment has a prominent ciliary band that aids in locomotion. This species superficially resembles the juveniles of other dorvilleid genera. The prostomium has a single pair of eyes, chemosensory organs, and multiple prostomial glands. The jaw denticles are protrusible and serrated at the tips and are used to scrape organic debris from the substratum. Large mandibles also play a role in defense, mating, and brood care. Laboratory cultures show that *O. adherens* can create water currents with its cilia, allowing it to facultatively filter feed (Paavo et al. 2000).

**Life History.** *Ophryotrocha adherens* is a protandric simultaneous hermaphrodite, as detailed by Paavo et al. (2000). There is courtship behavior before mating when a functional male attaches to a female. The female constructs a mucus cocoon integrating fecal material and detritus. The time between egg deposition in the cocoons to the first spawning is 18–26 days at HOC. A brief larval period ensures that enough young recruit to and remain near the parent population. Individuals can produce several broods of approximately 60 eggs over their typically 20-week life span. High abundances can be attained over a few weeks.

**Nereididae**  
*Neanthes arenaceodentata* (Moore, 1903)

These are some of the largest worms (5–12 mm long) found in sandy sediments near the outfalls. Adults have 29–54 segments and well-developed parapodia typical of active, motile polychaetes. The head bears two pairs of eyes, a pair of palps, and an eversible muscular pharynx with jaws for grasping food and numerous conical teeth that aid ingestion. *Neanthes arenaceodentata* is an omnivore that feeds on particles and debris of various sizes. The jaws also play a role in courtship and aggression (Reish 1957).

**Life History.** *Neanthes arenaceodentata* has been cultured in laboratory aquariums and used as bioassay organisms (Reish 1957, 1973). Paired worms produce exclusively benthic juveniles after 100 days (first spawning) that remain within the parent population. An omnivorous feeding mode that can directly (scavenging) or indirectly (predation) utilize sediment organics, brood protection, absence of a pelagic larval stage, and active behavior are traits that make this a successful species in food-rich sediments.

Experiments by Pesch et al. (1987) suggested that *N. arenaceodentata* is sensitive to intraspecific density as expressed in the size of paired mature females and their time to first spawning. Those authors appropriately cautioned assayists to consider intraspecific density in subacutely contaminated sediments. Pesch et al.'s highest experimental density was 486 individuals per square meter, whereas Hawaiian field densities of *N. arenaceodentata* were up to 30,000 individuals per square meter, with 7,054 individuals per square meter being median over several years. In contrast to measurements for laboratory-cultured specimens, our largest field specimens (+20 mm, 68 setigers) were typically from samples with higher densities. This may indicate that large specimens in low densities are under-sampled by our methods. It could also mean that the effects of density in our field conditions are strongly mitigated by environmental conditions near the outfall. Clearly, more work needs to be done at higher densities under various organic enrichment levels for bioassays to accurately detect sediment conditions that are not acutely polluted.

**Capitellidae**  
*Capitella capitata* (Fabricius, 1780)

Although different microscopic morphologies (such as setiger on which copulatory
setae appear) can be discerned in the laboratory, *Capitella capitata* and *Capitella* sp. A respond similarly in the environment and are collectively referred to here as the *C. capitata* complex. These are small worms (approximately 2–3 mm long) with reduced head structures and parapodia well adapted for burrowing. The jawless, protrusible proboscis engulfs sediment while burrowing. *Capitella capitata* is easily maintained in the laboratory in sediments with a high organic content. When present to the exclusion of other species it is considered a common indicator of excessive organic enrichment.

**LIFE HISTORY.** *Capitella capitata* males have spines used in mating, and females deposit fertilized eggs into a secreted mucus tube. Pelagic larvae may disperse to other locations or settle nearby. Adults and larvae are often used in bioassay experiments because of their association with sewage outfalls and tolerance of excessive organic enrichment.

**Syllidae**

*Pionosyllis heterocirrata* (Hartmann-Schröder, 1959)

These are small (2–3.5 mm) worms with three pairs of eyes; long, smooth antennae; an armed pharynx; a pair of palps; and well-developed parapodia. These active worms have muscular pharynges and appear to feed as detrital omnivores. This is one of the most abundant and widespread polychaete species around O'ahu, regularly representing 5 to 10% of all enumerated polychaetes at the ocean outfall.

**LIFE HISTORY.** Syllids may reproduce asexually by budding or sexually by several mechanisms. *Pionosyllis heterocirrata* has produced young in our laboratory within 3 weeks. Adults have been found with attached embryos that break off and grow as new recruits to the parent population. Eggs are often visible through the body wall. Larvae may disperse to new areas by a brief swimming period (Bailey-Brock 1996). Small size, frequent reproduction by sexual and asexual means, and a relatively short generation time lead to large numbers of *P. heterocirrata* in offshore sandy habitats. This study and work in other sandy O'ahu sediments (Bailey-Brock 1996, McCarthy 1996; J.H.B.-B., unpubl. data) suggest that the abundance of *P. heterocirrata* is largely unaffected by the input of organics.

**Sabellidae**

*Euchone* sp. B and *Augeneriella dubia* Hartmann-Schröder, 1965

Both species are small, tubicolous worms bearing a crown of branching radioles. Specimens of *Euchone* sp. B are typically 1.5 mm long, and *A. dubia* is usually twice as large. Sabellids are suspension feeders that use the ciliary surfaces of their crowns to trap and convey small particles to the mouth. They also sweep the sediment surface to collect appropriately sized particles. Grains within a particular size class are ingested, and larger grains may be incorporated into the tube as the worm grows or simply rejected.

**LIFE HISTORY.** *Euchone* sp. B and *A. dubia* may reproduce asexually by division within their tubes or sexually by producing pelagic larvae. Sabellids present an array of reproductive strategies from broadcast spawning to intratubular brooding. We have not observed reproductive traits in these species.

**RESULTS**

**Granulometry and Geochemical Findings**

Time and replicate averaged granulometric analyses (Figure 2) showed a larger component of fine grains (Φ 4 and finer) present at two far field sites (B1 and B2) compared with all other 70 m sites. Annual data show a significantly increasing proportion of particles Φ 4 and finer at stations B1 and B2 throughout the 1990–1998 study period (Table 1). The rate of increase of fine sediments at station B2 was approximately twice that of station B1. The few data points available from 1999–2000 data at sites D1 and D2 (50 m deep) continued this trend. Stations B3, B4, B5, Z, and B6 did not show any significant grain distribution trends, though annually averaged data of Φ 4 and finer grains at B4 suggest a slow accumulation. Averaged grain size distributions were uniform with medium and
coarse fractions in similar proportions at ZID stations B3, Z, B4, B5, and far field station B6.

Geochemical statistical results are summarized in Table 1. The total organic carbon (TOC) content of sediments ranged from 0.01 to 0.94% of dry sediment weight, with most measurements lying within the 0.02 to 0.3% range. Replicates within any station often varied by a factor of two to four. In 1992, TOC levels presented a wider range of values (0.01 to 0.95% dry wt.) spread uniformly among all stations. Overall, there was no significant difference in the measured TOC content of sediments from ZID sites when compared with those of far field sites in a two-sample t test (95% C.I., $t = -0.10$, $df = 127$, $P = 0.92$). Regression analysis found a slight significantly increasing trend in TOC levels at all study sites. Total Kjeldahl nitrogen (TKN) as a general nutrient indicator did not differ significantly between ZID...
TABLE 1
Geochemical Statistics of 1990–1998 Study Sites

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Equation</th>
<th>$r^2$</th>
<th>$P$</th>
<th>$n$</th>
<th>Range</th>
<th>Mean</th>
<th>SE Mean</th>
</tr>
</thead>
<tbody>
<tr>
<td>TOC (ZID)</td>
<td>$TOC = 0.325 - 0.0364x$</td>
<td>29.7%</td>
<td>0.000</td>
<td>76</td>
<td>0.01–0.94</td>
<td>0.136</td>
<td>0.0203</td>
</tr>
<tr>
<td>TOC (Non-ZID)</td>
<td>$TOC = 0.265 - 0.0261x$</td>
<td>16.0%</td>
<td>0.000</td>
<td>81</td>
<td>0.01–0.88</td>
<td>0.135</td>
<td>0.0189</td>
</tr>
<tr>
<td>TKN (ZID)</td>
<td>$TKN = 305 - 18.2x$</td>
<td>51.3%</td>
<td>0.000</td>
<td>76</td>
<td>28–523</td>
<td>210.70</td>
<td>7.72</td>
</tr>
<tr>
<td>TKN (Non-ZID)</td>
<td>$TKN = 261 - 13.3x$</td>
<td>39.8%</td>
<td>0.000</td>
<td>79</td>
<td>66–296</td>
<td>195.06</td>
<td>6.23</td>
</tr>
<tr>
<td>ORP (ZID)</td>
<td>$ORP = 34.8 + 18.4x$</td>
<td>45.9%</td>
<td>0.000</td>
<td>91</td>
<td>22–280</td>
<td>132.47</td>
<td>7.05</td>
</tr>
<tr>
<td>ORP (Non-ZID)</td>
<td>$ORP = 37.7 + 11.9x$</td>
<td>22.5%</td>
<td>0.000</td>
<td>90</td>
<td>3–260</td>
<td>97.01</td>
<td>6.85</td>
</tr>
<tr>
<td>B1 grains $\Phi \geq 4$</td>
<td>$% \geq 4 = 31.6 + 0.997x$</td>
<td>36.0%</td>
<td>0.001</td>
<td>27</td>
<td>27.3–46.5</td>
<td>36.55</td>
<td>0.842</td>
</tr>
<tr>
<td>B2 grains $\Phi \geq 4$</td>
<td>$% \geq 4 = 28.9 + 1.84x$</td>
<td>36.8%</td>
<td>0.000</td>
<td>29</td>
<td>17.9–55.9</td>
<td>38.34</td>
<td>1.46</td>
</tr>
</tbody>
</table>

$^a$ TOC, total organic carbon (% dry weight); TKN, total Kjeldahl nitrogen (mg N per wet kg sediment); ORP, oxidation-reduction potential (+mV); B1 and B2 grain fraction $\Phi \geq 4$ (% total dry weight).

and far field sites in a two-sample $t$ test (95% C.I., $t = 1.32$, df = 126, $P = 0.19$). TKN showed a significantly decreasing trend at all study sites over the 1990–1998 period. A two-sample $t$ test indicated that ZID sites produced significantly higher oxidation-reduction potential (ORP) values than non-ZID sites (95% C.I., $t = 3.37$, df = 169, $P = 0.0005$). Twenty percent of all ORP measurements fell below 50 mV; 12% at far field stations. With the observed variability between replicate measurements, it appears that aerobic conditions were maintained in superficial sediments throughout the study period at all sites. Regression analysis did not find a significant rise or fall of oil and grease levels over the course of the study period. A two-sample $t$ test did not indicate a significant difference in oil and grease levels between ZID and far field sites (95% C.I., $t = -0.50$, df = 145, $P = 0.61$). Oil and grease levels were routinely between 8 and 500 mg/kg of wet sediment except during 1992 when all samples ranged between 700 and 1400 mg/kg in similar site rankings.

Numerical Dominance Relative to the Outfall

Historical 70 m results show that *Ophryotrocha adherens* and *Euchone* sp. B had antithetical ZID/non-ZID distributions (Figures 3 and 4). *Ophryotrocha adherens* consistently and abundantly occupied ZID stations B3, B4, Z, and B5; *Euchone* sp. B specimens were rare at those sites, but numerous at the far field sites B1, B2, and B6. Stations B1 and B2 were distinct in having a much larger proportion of fine grains (Figure 2). *Euchone* sp. B abundance, as a percentage of polychaetes sampled, positively and significantly correlated with the increasing proportion of fine grains ($\Phi$ 4 and finer combined) at B2 over the 1990–1998 study period (correlation = 0.876, $P = 0.002$). A nonsignificant but direct correlation was present at stations B1, B4, and B6. *Euchone* sp. B typically composed 20–50% of all polychaetes collected at those sites in a given year. Although there were fewer 1999–2000 data, they appear similar. *Euchone* sp. B was more abundant at the 50 m sites, the depth most comparable with the historical 70 m data. *Euchone* sp. B abundance was depressed at the ZID sites where *O. adherens* was more abundant.

*Neanthes arenaceodentata* was numerically dominant at ZID site B3 from 1990 to 1992, composing 25 to 70% of all polychaetes, but was replaced by *Ophryotrocha adherens* in subsequent years. The general increase in *O. adherens* abundance at ZID stations was matched by a marked decrease in *N. arenaceodentata* over the same period. *Neanthes arenaceodentata* presents an interesting distribution among the sample sites. Most years produced specimens only from ZID sites. In 1990, 1992, and 1998 non-ZID sites produced very few small individuals, and ZID sites correspondingly produced exceedingly...
Ophryotrocha adherens Abundance

Figure 3. Percentage polychaete abundance of *Ophryotrocha adherens* during 70 m historical study. Nine columns are present for each station block representing 1990–1998 data from left to right. Vertical bars represent the range; closed circles denote the median; plus signs (+) locate outlying data points from individual replicates.

High numbers and our largest specimens. Abundance of *N. arenaceodentata* at non-ZID stations never exceeded 3% of collected polychaetes in any given year.

The *Capitella capitata* complex was consistently present at all 70 m sites except far field B6 (Figure 5). *Capitella* abundance fluctuated from year to year, but did so correspondingly at all sites. For example, relatively low abundance was measured at all sites in 1990. In 1992, their proportional abundance approximately doubled at all sites. This increase corresponded to an increase in the measured TOC and oil and grease levels. In 1996, we saw a decrease, again across all sites. The year 1996 presented uniformly low TOC and oil and grease levels. In each of the 10 years studied, *Capitella* composed a greater proportion of the polychaete collection at ZID site B3 than at any other site. Collectively, ZID sites did not produce a significantly higher proportion of *Capitella* than non-ZID sites, but far field stations always produced fewer *Capitella* individuals than ZID site B3.

The two sabellids, *Eucbene* sp. B and *Augeneriella dubia*, presented different distributions. *Eucbene* sp. B was extremely rare within ZID sites and numerous at far field sites B1, B2, and B6. In contrast, *A. dubia* composed 30–50% of the polychaetes at ZID sites B4 and Z in 1994. The sediment analysis that year correspondingly reported a coarser grain distribution than is typical for B4. *Augeneriella dubia* commonly represented 2–20% of the polychaetes collected at ZID station B5. *Eucbene* sp. B was abundant in the fine sediments of sites B1 and B2. Both species were present in moderate proportions at the far field B6 and were generally absent at ZID stations (with exceptions as already noted).
**Pionosyllis heterocirrata** was a commonly dominant species at all stations in all years. Abundance was stable and typically represented at least 5% of any given polychaete collection each year. Its numbers appeared depressed at sites B1 and B2 when compared with other 70 m samples. The 1999 and 2000 data showed it to be well represented at 20-, 50-, and 100-m isobaths.

Raw polychaete abundances were rarely seen as significantly different between ZID and far field stations in sequential between-year tests as reported annually (Swartz et al. 2001). To consider the 70 m data set as a whole, a two-sample t test was applied to all non-ZID versus an equivalent number (162) of randomly selected ZID samples. ZID samples produced significantly more worms than non-ZID samples (99% C.I., \( P = 0.039 \)).

**Species Richness, Diversity (H'), Evenness (J), and K-Dominance**

When the 70 m data were considered as 162 observations of ZID versus non-ZID sites, species richness differences were significant. ZID sites produced an average of 18.93 species per sample, and non-ZID sites produced a mean of 22.40 species. This small difference was consistent enough to be significant (99% C.I., \( P = 0.0000 \)).

The ZID stations had a significantly higher \( H' \) value than non-ZID sites (99% C.I., \( P = 0.0000 \)). The mean \( H' \) was 1.989 for non-ZID sites and 2.338 for ZID sites. The ZID samples also presented a more even (mean \( J = 0.784 \)) grouping of polychaetes than non-ZID samples (mean \( J = 0.707 \)), which tested as significant (99% C.I., \( P = 0.0001 \)). Both of these values should be inter-

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**Figure 4.** Percentage polychaete abundance of *Euchone* sp. B during 70 m historical study. Nine columns are present for each station block representing 1990–1998 data from left to right. Vertical bars represent the range; closed circles denote the median; asterisks (*) locate outlying data points from individual replicates.
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Figure 5. Percentage polychaete abundance of *Capitella capitata* complex during 70 m historical study. Nine columns are present for each station block representing 1990–1998 data from left to right. Vertical bars represent the range; closed circles denote the median.

Interpreted with caution because of the inherent limitations of these statistics.

Figure 6 presents the K-dominance curves generated from 70 m data. Although the ZID curve is slightly depressed, both present similar convex profiles, with an intersection far to the right. Similar, nonintersecting shapes denote comparable assemblages amenable to diversity analyses. A polluted site is expected to present a distinctly different (often concave) profile when compared with a presumed unimpacted site presenting a convex curve. Lambshead et al. (1983) have established that intersecting K-dominance curves as calculated here indicate that assumptions made by the $H'$ and $J$ statistics have been violated.

MDS Community Analyses

Anosim analyses were run before viewing MDS ordinations to avoid a priori grouping assumptions. One-way tests comparing ZID versus non-ZID locations strongly supported a distinction with an $R$ value of 0.414 for 1990–1993 data and an $R$ value of 0.627 for 1994–1998 data. An $R$ value of 0.45 corresponds to an unlikely event, with a probability of less than 1 in 1000 if there is no difference between ZID and non-ZID samples. The statistical significance of this type of analysis is necessarily low and can only ever reach 10% (Clarke and Warwick 1994). Both data sets returned a significance value of 0.0%. Two-way nested Anosim comparisons for distinguishing individual sites (years pooled) produced an $R$ value of 0.426 (significance level 5.7%) for 1990–1993 and an $R$ value of 0.963 (significance level 2.9%) for 1994–1998. These results illustrate each site’s independence, but support a distinction between ZID and non-ZID stations.

MDS ordinations for 70 m from 1990–
1993 (Figure 7) and from 1994–1998 (Figure 8) clearly show a separation between ZID (Z) and non-ZID (N) samples throughout the study period. When individual stations (years pooled) are labeled, rather than simply non-ZID versus ZID samples, groupings within each of the two major clusters indicate the variability of each site (Figures 9 and 10). In 1994–1998 data (Figure 10), stations B1 and B2 appear clustered but highly variable. Samples from B5 and B6 nearly span the ZID/non-ZID condition. Samples from B3, B4, and Z define the ZID condition. Comparisons using only the 10 species numerically dominant in the ZID or non-ZID samples during the 70 m study period produced nearly identical results with a distinct separation between ZID and non-ZID groups. However, the dispersion of samples within each major group was more diffuse. This led us to the interpretation that our dominant species were the major ordination influences; the presence or absence of rare species provided finer resolution within the major clusters.

Samples collected during the 1999 and 2000 study period produced only six polychaete species new to the Sand Island Wastewater Outfall monitoring program. Figure 11 presents the MDS ordination of the 45 samples collected in 1999 along three depth transects (site series CI–6, DI–6, and EI–6) as compared with 79 randomly selected ZID samples from the 1990–1998 study. Figure 12 similarly presents the 2000 data with the same historical 70 m ZID sam-

**Figure 6.** K-dominance curve created with 162 randomly selected samples (92 ZID and 70 non-ZID) from 1990–1998 70 m study.
FIGURE 7. Primer MDS ordination (stress = 0.24) of 1990–1993, 70 m data. N, non-ZID; Z, ZID.

DISCUSSION AND CONCLUSIONS

The Sand Island Wastewater Outfall typically discharges 2.6 to 3.4 m³/sec of advanced primary effluent through 285 ports located along nearly 2 km of diffuser at depths ranging from 67 to 72 m. Although the outfall is obviously releasing nutrients into the environment in the form of dissolved and particulate organic carbon, it does not appear that a measurably substantial amount is deposited as detritus directly available to benthic epifauna and infauna. Dollar, as summarized in Nelson et al. (1992a,b), reported that little (<1%), if any, particulate matter ever reaches the sediment surface. Roberts (1996) measured bottom currents in the immediate vicinity of historical site B5 as predominantly westward and typically ranging from 0 to 30 cm/sec, rarely reaching 50 cm/sec. The geochemical parameters reported here indicate that very little benthic accumulation of effluent is occurring, likely due to the local water flow regime combined with the effluent’s low salinity.

Although organic enrichment is often presumed to be the primary influence on adjacent benthic communities, municipal effluents and the physical structure of outfalls in nutrient-poor tropical waters may present additional confounding factors. Riddle et al. (1990) reported that although there is a pau-
city of data on benthic productivity in tropical sediments, studies by Ansell and Parulekar found that tropical biomass often had a turnover rate up to 10 times that of temperate systems. The physical structure of the outfall and its capstones acts as an artificial reef that supports a greater standing stock of fish and sessile invertebrates (Brock 2001) than the surrounding sandy environment. Posey and Ambrose (1994) and Davis et al. (1982) demonstrated that fish aggregated by artificial structures or isolated hard substrata have a significant effect on shaping benthic communities within a 75- to 140-m radius.

Nine years of monitoring efforts at 70 m identified three polychaetes, Ophryotrocha adherens, Neanthes arenaceodentata, and the Capitella capitata complex, closely associated with sediments immediately surrounding the Sand Island Wastewater Outfall. Neanthes arenaceodentata and the C. capitata complex are recognized internationally as pollution indicators. Capitella capitata tolerates higher sediment organic levels, including hydrocarbons, and anaerobic conditions more than most polychaete species, as demonstrated experimentally by Bridges et al. (1994). The decline of N. arenaceodentata was one of the most noticeable changes in the benthic community at the Sand Island Wastewater Outfall. It seems unlikely that the smaller worm, O. adherens, actively displaced N. arenaceodentata at the ZID. Both species have life history traits that favor juvenile recruitment into the parent population, leading to rapid population increase in the immediate vicinity.
and reduced dispersal to new locations. An explanation for this change may be that *O. adherens* reproduces faster and reaches higher abundances on a smaller spatial scale (typically 8000 to 20,000 individuals per square meter) than *N. arenaceodentata*. The nereidid life cycle includes a longer generation time, a large size, and a high level of activity, which suggests that this worm may not have been sampled adequately with the 7.6-cm-diameter corer. Both species share the characteristics of an active lifestyle: young that are retained in the parent population close to food resources, and an omnivorous feeding mode. However, the smaller species numerically dominated the ZID benthos. We suggest that *N. arenaceodentata* abundance has dramatically declined from ubiquity to almost complete local extinction due to an unknown cause and the smaller *O. adherens* abundance rose in the absence of *N. arenaceodentata* competition. Three years in which *N. arenaceodentata* was detected outside the ZID corresponded with extremely high ZID abundances. We believe this supports the supposition that conditions surrounding the outfall enhance *N. arenaceodentata* reproduction and that small specimens recorded from the ZID sites represent dispersed young. These young seem to be poor colonists of distant sediments.

The well-known indicator *Capitella capitata* has been present at the outfall since 1986 and continues to exist in moderate abundance. It has also been found at fish farms using net cage culture in Japan and off O'ahu (Tsutsumi 1987, Tsutsumi et al. 1990, Bybee and
Bailey-Brock in press), near sewage outfalls (Nelson et al. 1994), and outside the mouth of the eutrophic Ala Wai Canal (McCarthy 1996, McCarthy et al. 2000). The abundance of the *C. capitata* complex did not appear to be affected by the grain size distributions encountered in this study. These polychaetes were consistently present in higher numbers near the outfall and increased at other stations during periods of elevated TOC and oil and grease levels.

*Augeneriella dubia* is a ZID-associated species with feeding, reproduction, and life history traits similar to those of *Euchone* sp. B. The former appears to do particularly well in coarse sands, and the latter thrives in finer sediments. *Euchone* sp. B is a common inhabitant of control stations, but it has not played an important role in communities immediately adjacent to the outfall. Absence at the ZID points to the lack of suitable habitat for this species.

*Pionosyllis heterocirrata* has been consistently abundant throughout this study. Other studies (McCarthy 1996, Bybee and Bailey-Brock in press) have reported that it has been similarly abundant at other locations around O‘ahu. Any future reduction in abundance or absence at the ZID would be cause for concern, and in-depth monitoring would be advisable. This robust and active species is the most commonly encountered species in coral sands at these depths around O‘ahu shores.

Community analyses using abundance and diversity assessments applied to the Pearson-Rosenberg model will likely detect a large or
Species richness has been regularly calculated and compared on a between-year and between-station basis during annual reporting. The results were rarely significant. No such pollution destructive to benthic assemblages seems to be present at the Sand Island Wastewater Outfall as determined by this study and by Swartz et al. (2001). Carefully considered species-level comparisons offer finer resolution by examining individual, biologically key components of the benthos adjacent to the outfall. In addition, MDS analyses offer an opportunity to examine the speciose polychaete assemblages as a whole, allowing a more finely resolved picture of the benthic environment. Any future samplings near the outfall should be placed in this historical data matrix to see if they appear more like a historical ZID or non-ZID station. In
this fashion the community transition and therefore the current "boundary" of impact might be delineated. It would be instructive to define and compare this "actual impact" boundary with the arbitrarily determined EPA boundary within the context of O'ahu's unique hydrogeographic environment to facilitate locally appropriate management decisions. Recent software developments will ensure that future assessments will be able to simultaneously analyze the full data set without resorting to the data reduction necessitated in this study.

Initial examinations of the 20, 50, and 100 m data (this study and Swartz et al. 2001) indicate that depth may be a stronger force in shaping the polychaete communities than proximity to the outfall for stations outside the ZID. Future sampling efforts should consider comparing individual yearly samples
with historical MDS ordinations to help delineate the outfall’s impact and to further examine the diversity of O‘ahu’s southern shore. Our analysis indicates that conditions adjacent to the outfall positively enhance polychaete abundance and species composition without substantially degrading the habitat. It is likely that any organic enrichment reaching the sediment from the outfall can be converted rapidly to faunal biomass. More detailed studies on the role of organic carbon and its measurement in sandy tropical sediments are required. The accumulation of fine sediment particles at control stations B1 and B2 does not appear to indicate an accumulation of outfall-associated particulates. Any potential enrichment effects appear to be confined to the ZID boundary. Future studies should include multiple factor analysis to further define the effects depth, proximity to the outfall (and other hard substrata), and sedimentary physical parameters have on shaping the benthic communities of the region. The sediment composition, specifically surrounding historic sites B1 and B2, should continue to be monitored with the goal of identifying the source of increasing fine particulates, perhaps utilizing recent advances in stable isotope analyses and carbohydrate measurement techniques. Dominant species monitoring and MDS analyses offer the potential of an “early warning system” of organic enrichment when compared with typical community level analyses in addition to providing fundamental insights into the forces shaping benthic communities and should form a regular part of future monitoring efforts.

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